

1 **Title:** Emergence and disruption of cooperativity in a denitrifying microbial community

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20 **ABSTRACT**

21  
22 Anthropogenic perturbations to the nitrogen cycle, primarily through use of synthetic fertilizers, is  
23 driving an unprecedented increase in the emission of nitrous oxide (N<sub>2</sub>O), a potent greenhouse  
24 gas and an ozone depleting substance, causing urgency in identifying the sources and sinks of  
25 N<sub>2</sub>O. Microbial denitrification is a primary contributor to biotic production of N<sub>2</sub>O in anoxic regions  
26 of soil, marine systems, and wastewater treatment facilities. Here, through comprehensive  
27 genome analysis, we show that pathway partitioning is a ubiquitous mechanism of complete  
28 denitrification within microbial communities. We have investigated mechanisms and  
29 consequences of process partitioning of denitrification through detailed physiological  
30 characterization and kinetic modeling of a synthetic community of *Rhodanobacter R12* and  
31 *Acidovorax 3H11*. We have discovered that these two bacterial isolates, from a heavily nitrate  
32 (NO<sub>3</sub><sup>-</sup>) contaminated superfund site, complete denitrification through the exchange of nitrite (NO<sub>2</sub><sup>-</sup>  
33 ) and nitric oxide (NO). The process partitioning of denitrification and other processes, including  
34 amino acid metabolism, contribute to increased cooperativity within this denitrifying community.  
35 We demonstrate that certain contexts, such as high NO<sub>3</sub><sup>-</sup>, cause unbalanced growth of community  
36 members, due to differences in their substrate utilization kinetics. The altered growth  
37 characteristics of community members drives accumulation of toxic NO<sub>2</sub><sup>-</sup>, which disrupts  
38 denitrification causing N<sub>2</sub>O off gassing.

39  
40 **INTRODUCTION**

41  
42 Since the development of the Haber-Bosch process, there has been an exponential increase in  
43 the global deposition of fixed nitrogen (N). This has largely resulted from the use of synthetic  
44 fertilizers<sup>1</sup>. Meeting the demands of a growing global population has contributed to a nearly ten-  
45 fold increase in synthetic fertilizer use from 1960 to 2013<sup>2</sup>. Recent estimates suggest  
46 anthropogenic contributions account for about half of reactive N flux on Earth<sup>2</sup>. As a result, the  
47 global N cycle has become increasingly perturbed, contributing to eutrophication of terrestrial and  
48 aquatic systems, global acidification, increasing atmospheric nitrous oxide (N<sub>2</sub>O) and  
49 stratospheric ozone loss<sup>1</sup>. The microbial processes of nitrification and denitrification are  
50 responsible for the bulk of N<sub>2</sub>O emissions in agricultural systems and wastewater treatment<sup>3-5</sup>.  
51 Thus, understanding how microbial processes, interactions, and environmental factors such as

52 resource concentration, pH, and metal availability influence the fate of N compounds, and the  
53 production of N<sub>2</sub>O in particular, is essential for developing a predictive understanding of the fate  
54 of different N species in natural and engineered systems<sup>6-9</sup>.

55 Denitrification is the major biological process that returns fixed N back to the atmosphere  
56 through reduction of N-oxides nitrate (NO<sub>3</sub><sup>-</sup>) and nitrite (NO<sub>2</sub><sup>-</sup>) to gaseous compounds nitric oxide  
57 (NO), N<sub>2</sub>O, and dinitrogen (N<sub>2</sub>). Therefore, it comprises a critical step in the global N cycle.  
58 Denitrification is also an anaerobic process, which provides an alternative means of generating  
59 energy when oxygen (O<sub>2</sub>) is not available. In complete denitrification, microorganisms reduce  
60 NO<sub>3</sub><sup>-</sup> to N<sub>2</sub> in a series of four reductive, energy-conserving steps. First, NO<sub>3</sub><sup>-</sup> is reduced to NO<sub>2</sub><sup>-</sup>  
61 by one of two dissimilatory nitrate reductases—the membrane-bound Nar complex or the  
62 periplasmic nitrate reductase, Nap. NO<sub>2</sub><sup>-</sup> is then reduced to NO by one of two structurally unrelated  
63 nitrite reductases that contain different prosthetic groups—the copper-containing NirK or the  
64 cytochrome cd1 NirS<sup>10</sup>. NO is then reduced to N<sub>2</sub>O by the nitric oxide reductase (NorB), and finally  
65 N<sub>2</sub>O is reduced to N<sub>2</sub> by the nitrous oxide reductase (NosZ), of which two distinct clades have  
66 been identified<sup>10,11</sup>.

67 Most studies have only examined complete denitrifiers<sup>12,13</sup> in monoculture to study this  
68 respiratory process (enzymology, kinetics, regulation, etc.). However, ongoing genomic and  
69 metagenomic characterization of isolates and natural communities has revealed, remarkably, that  
70 the genomes of most organisms that encode denitrification enzymes lack the complete pathway<sup>14</sup>.  
71 More commonly, the microbial communities in soil that drive denitrification are composed of a  
72 complex mixture of species that individually encode partial (rather than complete) denitrification  
73 pathways<sup>15-18</sup>. The environmental significance of organismal patchiness of genes in the pathway  
74 for denitrification is mostly unexplored but may be associated with increased efficiency and  
75 resiliency of denitrifying communities. Recent work has shown that pathway partitioning can  
76 decrease inter-enzyme competition within individual microbes and reduce the accumulation of  
77 toxic intermediates like NO<sub>2</sub><sup>-</sup> through their exchange, thus providing a potential selection

78 mechanism for pathway partitioning in denitrifying communities<sup>19</sup>. However, understanding how  
79 this generalizes when multiple soluble N intermediates (i.e.,  $\text{NO}_2^-$ , NO,  $\text{N}_2\text{O}$ ) are exchanged as  
80 well as how the environment selects for natural communities with specific partial pathway  
81 combinations remains unexplored. Additionally, because many soil and groundwater dwelling  
82 organisms are facultative anaerobes, the ability of these organisms to leverage multiple pathways  
83 and substrates for respiration may also contribute to denitrification pathway patchiness.

84 Here, we have investigated the role of pathway partitioning in a synthetic community  
85 (SynCom) composed of two isolates recovered from the same field site location at the Oak Ridge  
86 Reservation (ORR) (<https://enigma.lbl.gov>)<sup>20</sup> and representative of naturally occurring  
87 populations capable of exchange of pathway intermediates. This two-organism SynCom served  
88 as a model for examining the roles of biotic and abiotic factors controlling  $\text{N}_2\text{O}$  emissions in the  
89 low-oxygen and anaerobic zones of the subsurface environment of the ORR, which have high  
90 concentrations of  $\text{NO}_3^-$  due to past nuclear waste disposal and evidence of high levels of  
91 denitrification activity<sup>20,21</sup>. Using this SynCom as a model system, we have discovered how  
92 process partitioning of denitrification and exchange of multiple N intermediates improved  
93 community growth characteristics via changes in pathway kinetics (e.g., increased rate of  $\text{NO}_3^-$   
94 reduction, removal of  $\text{NO}_2^-$ , and exchange of  $\text{N}_2\text{O}$ ). We also show the importance of community  
95 composition in determining the physiology and growth dynamics of each species and how  
96 changes in environmental conditions can break the collaboration between the two members,  
97 leading to  $\text{N}_2\text{O}$  emissions.

98

## 99 **RESULTS**

100

101 *Denitrification pathway is partitioned across members of microbial communities at the ORR.*

102 We sought to understand the structure and activity of communities responsible for denitrification  
103 at the ORR and their relation to  $\text{N}_2\text{O}$  production. We first investigated the distribution of

104 denitrification pathway enzymes within the genomes of 252 isolates obtained from groundwater  
105 wells of various depths distributed across 3 major areas of the ORR<sup>20-23</sup>. The sites spanned areas  
106 of high, medium, and low average groundwater NO<sub>3</sub><sup>-</sup> concentration (between 0-250 mM NO<sub>3</sub><sup>-</sup>).  
107 For the pathway analysis we considered the genes Nar (or Nap), Nir, Nor, and Nos. Collectively,  
108 the reactions encoded by these genes convert NO<sub>3</sub><sup>-</sup> to N<sub>2</sub> (Fig. 1A). Genomic analysis revealed  
109 that the majority (~71%) of isolates were incomplete denitrifiers (i.e., missing one or more genes  
110 of the pathway). Of these, 40% likely contributed to N<sub>2</sub>O production due to the presence of one  
111 or more genes encoding Nor but none encoding Nos. These conclusions were consistent with  
112 the pathway compositions of 4622 IMG soil isolates, a vast majority of which (90%) were also  
113 incomplete denitrifiers (Fig. 1A). Thus, the genomic analysis demonstrated that the denitrification  
114 pathway is largely partitioned across soil and ground water dwelling organisms and suggested  
115 that complete denitrification likely occurs through interactions among organisms with  
116 complementary capabilities within microbial communities.

117 To better understand the role of pathway partitioning in denitrifying communities, we  
118 assembled the field isolates *Rhodanobacter* sp. FW510-R12 (R12) and *Acidovorax* sp. GW101-  
119 3H11 (3H11) into a SynCom capable of complete denitrification through pathway intermediate  
120 exchange. The selection of organisms in this pairing was based on relationships among genus  
121 abundance, groundwater chemistry (e.g., NO<sub>3</sub><sup>-</sup> and N<sub>2</sub>O concentrations), and isolate co-  
122 occurrence determined through genomic and 16S rRNA profiling across the field site (Fig. 1B and  
123 C). We observed significant correlation between NO<sub>3</sub><sup>-</sup> and N<sub>2</sub>O concentrations in the ORR  
124 groundwater (Fig. 1C), consistent with previous observations that high NO<sub>3</sub><sup>-</sup> can lead to N<sub>2</sub>O  
125 emissions<sup>3,20,22</sup>. We also found that the abundance of *Rhodanobacter* spp. was correlated with  
126 both NO<sub>3</sub><sup>-</sup> and N<sub>2</sub>O levels at the field site (Fig. 1C), which was consistent with the reported high  
127 abundance of this genus in regions of high NO<sub>3</sub><sup>-</sup> and heavy metal contamination in the ORR<sup>8,23</sup>.  
128 While the abundance of *Acidovorax* spp. was weakly associated with NO<sub>3</sub><sup>-</sup> and N<sub>2</sub>O

129 concentrations (Fig. 1C), it was positively correlated with *Rhodanobacter spp.* (Fig. 1C) in non-  
130 acidic groundwater ( $\text{pH} \geq 7$ ). Flux balance analysis using constraints-based metabolic models  
131 indicated that a SynCom of R12 and 3H11 could perform complete denitrification through the  
132 exchange of  $\text{NO}_2^-$  and  $\text{N}_2\text{O}$  (Fig. 1D). Consistent with model predictions, 3H11 and R12 were both  
133 capable of growth on anaerobic minimal medium containing  $\text{NO}_3^-$  and acetate as the primary  
134 electron acceptor and donor pair, respectively. Additionally, consistent with the absence of a  
135 genome-encoded  $\text{NO}_2^-$  reductase (Nir), 3H11 consumed  $\text{NO}_3^-$  and accumulated  $\text{NO}_2^-$  (Fig. 1E).  
136 By contrast, R12 consumed  $\text{NO}_3^-$ , and accumulated  $\text{N}_2\text{O}$  with no transient accumulation of  $\text{NO}_2^-$ ,  
137 implying that the rate of  $\text{NO}_2^-$  reduction was equal to or greater than the rate of  $\text{NO}_3^-$  reduction  
138 (Fig. 1E).

139 The SynCom growth rate was significantly greater than that of monocultures of either of  
140 the two isolates (Fig. 1F). Additionally, the rate of  $\text{NO}_3^-$  reduction was increased relative to the  
141 individual monocultures, followed by a transient accumulation of  $\text{NO}_2^-$ , which was fully reduced to  
142  $\text{N}_2$  gas based on the lack of  $\text{N}_2\text{O}$  and ammonia accumulation (Fig. 1E). In depth growth  
143 characterization demonstrated that 3H11 and R12 complemented each other to perform complete  
144 denitrification, which manifested in synergistic improvement in overall growth dynamics relative  
145 to monocultures of the two organisms.

146

147 *Exchange of intermediates in the denitrification pathway largely explain enhanced growth*  
148 *phenotype of the SynCom.*

149 We hypothesized that synergistic improvement in rate of  $\text{NO}_3^-$  reduction and overall growth  
150 characteristics of the SynCom were likely due to R12-mediated rescue of  $\text{NO}_2^-$ -mediated growth  
151 inhibition of 3H11, which in turn reduced  $\text{N}_2\text{O}$  to  $\text{N}_2$ . To test this hypothesis, we characterized the  
152 growth kinetics of each monoculture in media supplemented with varied concentrations of  $\text{NO}_3^-$ ,

153  $\text{NO}_2^-$ , and  $\text{N}_2\text{O}$  (Fig. S1) and used that data to develop kinetic models of growth dynamics of each  
154 isolate and the SynCom. The growth rate of 3H11 was significantly greater than R12 across a  
155 wide range of  $\text{NO}_3^-$  (Fig. 2A), achieving its maximum growth rate at a much lower concentration  
156 ( $K_{S_{3H11}} \ll K_{S_{R12}}$ , Welch's t-test  $t=26.1$ ,  $p < 10^{-6}$ , Fig. S1F). While a low concentration of  $\text{NO}_2^-$   
157 supported a significantly higher growth rate of R12 relative to its growth rate on  $\text{NO}_3^-$  (10-fold  
158 greater for 1-5 mM substrate, Fig. 2A), growth was reduced at higher concentrations ( $K_{i_{R12}}=13.23$   
159 mM, Fig. S1F) and completely inhibited at 10 mM  $\text{NO}_2^-$  (Fig. S1B and D). Nitrite also had a strong  
160 inhibitory effect on the growth rate and biomass yield of 3H11, with complete inhibition above 5  
161 mM  $\text{NO}_2^-$  in media containing 10 mM  $\text{NO}_3^-$  ( $K_{i_{3H11}}=9.11$  mM; Figs. 2A and S1B, D and F). Notably,  
162 R12 appeared to not conserve energy from the reduction of  $\text{NO}_3^-$ , since yield and growth rates  
163 were identical on  $\text{NO}_3^-$  and  $\text{NO}_2^-$  ( $\text{slope}_{\text{NO}_3}=0.033 \pm 0.003$ ,  $\text{slope}_{\text{NO}_2}=0.032 \pm 0.006$  for ordinary  
164 least squares fits to linear portion of substrate versus maximum OD600; substrate concentrations  
165 1-10 mM; Welch's t-test  $t=1.1$ ,  $p=0.27$ ). Thus, R12 growth was modeled as entirely dependent on  
166 the reduction of  $\text{NO}_2^-$  and NO. Finally, while 3H11 was able to accumulate biomass by reducing  
167  $\text{N}_2\text{O}$ , its growth rate on  $\text{N}_2\text{O}$  was constant and not dose-dependent (2-30 mM;  $K_{S_{3H11}} \approx 0$ , Fig. 2A  
168 and S1D and F).

169 Kinetic models were developed based on a modified Monod framework, integrating logistic  
170 representation of carrying capacity into equations describing growth kinetics as a function of  
171 metabolite concentrations (Fig. 2B; see Methods for more details)<sup>24</sup>. Equations in the model were  
172 parameterized using maximum growth rates as a function of substrate, half velocity constants,  
173 and carrying capacities extracted from the growth data using Logistic or Monod fits (Figs. 2A and  
174 S1D and F). Linear models were developed to capture relationships between  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , and  $\text{N}_2\text{O}$   
175 concentration and maximum OD600 (Fig. S1E) and used in simulations to predict monoculture  
176 and SynCom growth kinetics in new experiments. The simulation was optimized to achieve better  
177 accuracy by using initial biomass as a free model parameter to effectively adjust lag phase. An

178 initial model formulation for the reduction of  $\text{NO}_3^-$  to  $\text{N}_2$ , combining the reduction of  $\text{NO}_2^-$  and  $\text{NO}$   
179 into a single reaction (i.e., reduction of  $\text{NO}_2^-$  to  $\text{N}_2\text{O}$ ), yielded reasonably accurate prediction of  
180 metabolite turnover and growth kinetics for R12 (Fig. S2B). However, the period of maximum  
181 growth was overestimated for 3H11 and the carrying capacity (max OD600) achieved by R12 was  
182 not well represented. Furthermore, prediction of SynCom dynamics was quite poor (Fig. S2B).  
183 While the introduction of  $\text{NO}_2^-$  inhibition improved predicted growth dynamics of monocultures and  
184 SynCom, overall model accuracy was still poor (Fig. 2C, S2C and D). The period of exponential  
185 growth was overestimated and  $\text{NO}_2^-$  turnover was not well predicted.

186         These results challenged our model assumptions and forced us to reconsider the scheme  
187 of metabolite exchange. The most plausible variation was to include exchange of  $\text{NO}$ , assuming  
188 that a significant fraction of the  $\text{NO}$  produced by R12 is available for reduction by 3H11. This  
189 scenario would be possible if the affinity and rate of  $\text{NO}$  reduction was considerably higher for  
190 3H11 than R12, as was the case for  $\text{NO}_3^-$ . The contribution of the reduction of  $\text{NO}_2^-$  to  $\text{N}_2\text{O}$  was  
191 subsequently doubled for 3H11 (assuming a similar contribution to its growth) and halved for R12,  
192 which was estimated from the energetics of these processes due to absence of measured rates  
193 of  $\text{NO}$  reduction. This assumption significantly improved predicted growth dynamics of the  
194 SynCom and offered explanation for the dominance of 3H11 in the community, in contrast to its  
195 relatively poor growth in monoculture (Fig. 2C, S2C). Thus, kinetic modeling was essential to  
196 revealing the metabolite exchange governing SynCom growth dynamics.

197

198 *Community growth drives global changes in transcriptional states of 3H11 and R12.*

199 In order to characterize the impact of paired growth on organismal physiology, we compared  
200 temporal changes in transcriptomes of each isolate grown as a monoculture and as a SynCom  
201 member. Transcriptomes were longitudinally profiled at four time points that corresponded with  
202 early-, mid-, and late-phases of growth for all cultures with the exception of the 3H11 monoculture,  
203 which was profiled at two time points because of its fast growth rate and limited biomass (Fig. 3A



204 and B). Using DESeq2<sup>25</sup> to identify differentially expressed genes (DEGs), we discovered that  
205 321 of 4,670 genes in 3H11 and 480 of 3,493 genes in R12 were differentially regulated across  
206 different growth phases in the mono- and co-culture contexts (Fig. 3C, Supplementary Table 1).  
207 K-means clustering and pathway ontology enrichment analysis (Fig. 3D) identified significant  
208 expression changes in a few metabolic pathways, including denitrification, amino acid  
209 metabolism, and central carbon metabolism (Fig. 3D and E).

210 Expression patterns supported the hypothesis that 3H11 is the primary reducer of NO and  
211 N<sub>2</sub>O when paired with R12 (Fig 3E). Both 3H11 *norB* and *nosZ* were significantly upregulated in  
212 the SynCom context (Welch's t-test  $t=-11.8$ ,  $p<10^{-6}$  for *norB*;  $t=-4.5$ ;  $p=0.003$  for *nosZ*). In addition  
213 to its overall higher average expression in the SynCom context, *norB* was further upregulated  
214 during NO<sub>2</sub><sup>-</sup> reduction, reaching maximum levels in late-log phase (T3). Transcript levels of 3H11  
215 *nosZ* increased from early- (T1) to mid-log phase (T2) in the SynCom context and decreased  
216 thereafter, mirroring corresponding changes in NO<sub>2</sub><sup>-</sup> levels (Fig. S3A). In contrast, transcript  
217 levels of all R12 denitrification genes were relatively stable across conditions (Fig. 3E), except for  
218 *nirK*, which was differentially expressed in the SynCom during the late-log phase of growth (Fig.  
219 S3B). Transcript level of *nirK* increased in the SynCom relative to monoculture growth during the  
220 period of active NO<sub>2</sub><sup>-</sup> reduction ( $\log_2$  fold change=2.62,  $p<10^{-6}$  for comparison of T4 *nirK*  
221 expression; Fig. S3B). Similarly, R12 *narG* was achieved maximum expression levels earlier (95  
222 hrs) in the SynCom relative to the monoculture context (119 hrs, Fig. 3E, Fig. S3B), which  
223 coincided with the timing of NO<sub>2</sub><sup>-</sup> accumulation in each of these contexts. This observation was  
224 consistent with a higher maximum growth rate and shorter lag phase of R12 monocultures during  
225 growth on NO<sub>2</sub><sup>-</sup> relative to NO<sub>3</sub><sup>-</sup> (Fig. 2A, S1A, S1C) and the increased rate of NO<sub>3</sub><sup>-</sup> reduction by  
226 the SynCom relative to monocultures (Fig. S2A).

227 Transcriptional changes across multiple pathways suggested additional mechanisms of  
228 competition and cooperation between R12 and 3H11 as a SynCom. For example, expression  
229 patterns for 6 of 34 tricarboxylic acid (TCA) cycle genes in 3H11 and 4 of 33 genes in R12

230 coincided with the timing of  $\text{NO}_3^-$  and  $\text{NO}/\text{N}_2\text{O}$  reduction across growth contexts (Fig. 3E).  
231 Expression of TCA cycle genes in 3H11 coincided with reduction of  $\text{NO}_3^-$  (T1 in both monoculture  
232 and SynCom contexts) and with  $\text{NO}_2^-$  (T4 in SynCom context) (Fig. S3A). Unlike 3H11, expression  
233 of TCA cycle genes in R12 were lower in the initial time points, when  $\text{NO}_3^-$  was reduced, and  
234 increased in later phases of growth, when  $\text{NO}_2^-$  and  $\text{N}_2\text{O}$  were being reduced (e.g., R12\_3138,  
235 *fumA*, downregulated in T1,  $\log_2$  fold change=-2.14,  $p < 10^{-6}$  and upregulated in T4,  $\log_2$  fold  
236 change=3.06,  $p < 10^{-6}$ ; Fig S3B). Additionally, 38 motility-associated genes (e.g., *cheA-Z*, *flgA-M*,  
237 *fliD-Q*, and *pilA*) were upregulated in R12 in the SynCom context, which may be a mechanism for  
238 interaction with 3H11 or competition for  $\text{NO}_3^-$  (Fig. 3E). Interspecies exchange of amino acids in  
239 the SynCom was suggested by the differential regulation of amino acid metabolism genes (74  
240 total in 3H11 and R12). While both species upregulated biosynthesis of branched-chain amino  
241 acids (BCAAs) in the SynCom, there was coordinated expression of genes for leucine efflux in  
242 3H11 and upregulation of BCAA degradation genes in R12 (Supplementary Table 1). The  
243 coordinated exchange of individual amino acids suggested by the expression data was also  
244 consistent with a SynCom metabolic model predicting a 2-3% increase in yield with exchange of  
245 individual BCAAs (Fig. S3C). Collectively, the gene expression analysis provided support for the  
246 hypothesis that 3H11 is the primary reducer of  $\text{NO}$  and  $\text{N}_2\text{O}$  and identified potentially additional  
247 mechanisms of metabolic interplay contributing to the improved growth characteristics of the  
248 SynCom.

249

250 *Proteomics analysis of the SynCom supports the dominant role of 3H11 across varied  $\text{NO}_3^-$*   
251 *concentrations and reveals global physiological adaptations.*

252 We investigated the consequence of environmental context on the composition and functional  
253 dynamics of the SynCom, by performing mass spectrometry-based proteomic profiling across  
254 varying initial co-culture conditions, including  $\text{NO}_3^-$  concentrations (1-40 mM  $\text{NO}_3^-$ , 50% 3H11)  
255 and community compositions (10 mM  $\text{NO}_3^-$ , 5-98% 3H11) (Fig. S4). Remarkably, regardless of

256 the ratio of the two organisms in the inoculum, abundance of unique peptides (Fig. S4C) revealed  
257 that at steady state the SynCom composition converged to ~65% 3H11 and ~35% R12 (Fig. 4A),  
258 which was consistent with the kinetic model-predicted dominant role of 3H11 (Fig. 2C)<sup>26</sup>.

259 Altogether, 36% of the variance in abundance profiles of 2,401 SynCom proteins was  
260 captured by variation in NO<sub>3</sub><sup>-</sup> concentration of the growth medium ( $r=-0.94$ ,  $p<10^{-6}$ ; Fig. 4B). In  
261 fact, abundance changes in nearly half of the proteome (1,105 proteins) were significantly  
262 correlated with NO<sub>3</sub><sup>-</sup> concentration ( $r\leq 0.5$  or  $r\geq 0.5$  and  $q<0.05$ ; Fig. 4C). Globally, protein  
263 abundance changes were also significantly influenced by initial community composition, which  
264 showed moderate correlations with PC1 and PC2 ( $r=0.58$ ,  $p<10^{-6}$ ;  $r=0.32$ ,  $p=0.011$  for PC1 and  
265 PC2 respectively; Fig. 4B). Changes in abundance of 206 proteins were significantly correlated  
266 with community composition ( $r\leq 0.5$  or  $r\geq 0.5$  and  $q<0.05$ ). Broadly, most proteins correlated with  
267 NO<sub>3</sub><sup>-</sup> concentration and community compositions were from specific pathways and processes,  
268 including energy metabolism, central carbon metabolism, and amino acid metabolism (Fig. 4C).

269 Abundance patterns in denitrification pathway enzymes suggested that both R12 and  
270 3H11 have distinct regulatory schemes that benefit overall growth characteristics of the SynCom  
271 across a range of NO<sub>3</sub><sup>-</sup> concentrations (Fig. 4D). For instance, the inverse correlation of 3H11  
272 NarG protein abundance and NO<sub>3</sub><sup>-</sup> levels suggests that regulation of this enzyme is highly  
273 sensitive to repression by NO<sub>2</sub><sup>-</sup>. By contrast, expression of R12 NarG was upregulated by low  
274 concentrations of NO<sub>3</sub><sup>-</sup> (<10mM), but was repressed at concentrations >10mM, also likely due to  
275 NO<sub>2</sub><sup>-</sup> accumulation. The abundance of R12 NirK was significantly correlated with NO<sub>3</sub><sup>-</sup>  
276 concentration, suggesting that its expression was activated by and proportional to NO<sub>2</sub><sup>-</sup> levels.  
277 Together these patterns suggest regulatory mechanisms that balance activities of NarG and NirK  
278 to limit the accumulation of NO<sub>2</sub><sup>-</sup> in the SynCom. Interestingly, trends in NorB were opposite for  
279 R12 and 3H11. Abundance of NorB increased for 3H11 and decreased for R12 as a function of  
280 NO<sub>3</sub><sup>-</sup>. Additionally, 3H11 NosZ abundance decreased with NO<sub>3</sub><sup>-</sup> (Fig. 4D). Less pronounced but

281 similar trends were observed for NarG and NirK across variation in community composition (Fig.  
282 4E). The decrease in NorB and NosZ with increase in the initial abundance of 3H11 (Fig. 4E) is  
283 likely due to NO<sub>2</sub><sup>-</sup> toxicity, since the rate of production of NO<sub>2</sub><sup>-</sup> by 3H11 exceeds its reduction by  
284 R12. Thus, correspondence in trends between NarG and NirK across conditions suggested that  
285 NO<sub>2</sub><sup>-</sup> accumulation resulted from an increase in NO<sub>3</sub><sup>-</sup> concentration as well as an increase in the  
286 proportion of 3H11 in the initial composition of the SynCom.

287 Protein abundance patterns across multiple pathways showed opposite trends between  
288 3H11 and R12 (Supplementary Table 2). For instance, whereas 40 of the 67 central carbon  
289 metabolism enzymes in R12 that were assayed were negatively correlated with NO<sub>3</sub><sup>-</sup>, 68 of 87  
290 enzymes in the corresponding pathways in 3H11 were positively correlated with NO<sub>3</sub><sup>-</sup> (Fig. 4F,  
291 Supplementary Table 2). Similar anti-correlated trends in protein abundance changes in 3H11  
292 and R12 were also observed with respect to initial community composition (e.g., proteins of TCA  
293 and glyoxylate cycles). Collectively these trends in protein abundance changes suggest that 3H11  
294 outcompeted R12 for substrates (i.e., acetate, NO<sub>3</sub><sup>-</sup>, NO etc.) with competition for substrates  
295 becoming more pronounced with increase in relative proportion of 3H11 in the inoculum, as well  
296 as with higher concentrations of NO<sub>3</sub><sup>-</sup> (Fig. 4G). Alternatively, these trends may suggest that key  
297 metabolic processes, such as amino acid biosynthesis pathways (Supplementary Table 2), were  
298 partitioned across the two organisms, complementing each other's nutritional needs and  
299 capabilities. For instance, while four enzymes of methionine biosynthesis in R12 were  
300 downregulated with increasing NO<sub>3</sub><sup>-</sup> concentration, a methionine importer, MetQ, was upregulated  
301 ( $r=0.7$ ,  $p<10^{-6}$ ). This regulation in R12 was likely mediated at the transcriptional level by MetJ, a  
302 putative repressor of methionine biosynthesis, which was upregulated with increased NO<sub>3</sub><sup>-</sup>  
303 ( $r=0.53$ ,  $p=5.6\times 10^{-5}$ ). By contrast, enzymes of methionine biosynthesis in 3H11 were upregulated  
304 with increased NO<sub>3</sub><sup>-</sup> concentration ( $r>0.5$ ,  $p<10^{-6}$  for MetE, MetH and MetY). These trends  
305 suggested that methionine biosynthesis in 3H11 was complementing the nutritional need of this  
306 amino acid in R12. The NO<sub>3</sub><sup>-</sup>-induced shifts in abundance of enzymes of amino acid metabolism

307 in both R12 and 3H11 was consistent with corresponding mRNA level changes indicating that  
308 much of this regulation was mediated at the transcriptional level. Thus, the proteomics analysis  
309 supported conclusions from transcriptomics analysis regarding the role of process partitioning  
310 and interplay of varied metabolic processes, including denitrification, central carbon metabolism,  
311 and amino acid metabolism, in enhancing growth characteristics of the SynCom.

312

313 *Nitrite accumulation drives nitrous oxide emissions across variation in initial nitrate*  
314 *concentration and community composition.*

315 We investigated how variation in abiotic factors (specifically,  $\text{NO}_3^-$  concentration) and biotic  
316 properties (specifically, community composition) impact  $\text{N}_2\text{O}$  emissions. We measured the steady  
317 state levels of  $\text{N}_2\text{O}$  in SynCom cultures across a range of initial  $\text{NO}_3^-$  concentrations (1 mM vs 40  
318 mM) and community composition (3H11:R12 ratio of 5%:95% vs 98%:2%). We expected that  
319 extremes in community composition and  $\text{NO}_3^-$  concentration might result in differences in  
320 community phenotype and  $\text{N}_2\text{O}$  production/consumption. We anticipated  $\text{N}_2\text{O}$  accumulation at  
321 higher concentrations of  $\text{NO}_3^-$ . At high  $\text{NO}_3^-$  concentrations acetate would be exhausted before  
322  $\text{NO}_3^-$  could be completely reduced to  $\text{N}_2$ , given that earlier steps in the pathway (e.g.,  $\text{NO}_3^-$  and  
323  $\text{NO}_2^-$  reduction) are prioritized by the SynCom. Consistent with this hypothesis,  $\text{N}_2\text{O}$  accumulation  
324 was proportional to initial  $\text{NO}_3^-$ . However, the observed accumulation of  $\text{N}_2\text{O}$  at 20 mM  $\text{NO}_3^-$  was  
325 well below the 30 mM threshold predicted by stoichiometry to become limiting. In addition, we  
326 were surprised to find that  $\text{N}_2\text{O}$  also increased with increasing proportions of 3H11 in the  
327 inoculum, despite its role as the primary  $\text{N}_2\text{O}$  reducer in the SynCom (Fig. 5A). Together these  
328 findings suggested a more complex interplay between 3H11 and R12 underlies the dynamics of  
329 denitrification by the SynCom.

330 To characterize possible mechanisms responsible for accumulation of  $\text{N}_2\text{O}$  by the  
331 SynCom, we examined associations between  $\text{N}_2\text{O}$  concentration and protein abundance across  
332 all conditions. We identified 13 proteins that were positively correlated with  $\text{N}_2\text{O}$  levels across

333  $\text{NO}_3^-$  concentrations and community composition ( $r \geq 0.5$  and  $q < 0.05$ ) and 12 proteins with  
334 abundance changes that were negatively correlated ( $r \leq -0.5$  and  $q < 0.05$ ). Proteins that were  
335 positively correlated with  $\text{N}_2\text{O}$  were largely from R12 (10 of 13) and of unknown function (6 of 10).  
336 The four R12 proteins with predicted functions were NirK, an iron transporter, Bfr (bacterioferritin),  
337 and Ahr (alcohol/geraniol dehydrogenase). The three 3H11 proteins that were positively  
338 correlated with  $\text{N}_2\text{O}$  were associated with central carbon and amino acid metabolism and included  
339 AceB, GlcB (malate synthase), Ddl (D-alanine-D-alanine ligase), and ilvH, ilvN (acetolactate  
340 synthase). In contrast, proteins with abundance changes negatively correlated with  $\text{N}_2\text{O}$  levels  
341 were primarily from 3H11 (10 of 12) and related to energy metabolism (7 of 10). Two of the seven  
342 energy metabolism proteins were cytochromes, while the rest were from the denitrification  
343 pathway, including two copies of NarG, NarH, NosZ, NosR (a transcriptional regulator of NosZ)  
344 and PhasZ (PHB depolymerase; Fig. 5B).

345 Using a multi-linear regression framework, we investigated the ability of protein  
346 abundances within specific pathways to predict observed final  $\text{N}_2\text{O}$  levels (Fig. 5C, see Methods  
347 for additional details). As expected, protein abundances within the N metabolism pathway had  
348 the best performance in predicting final  $\text{N}_2\text{O}$  levels (out of sample  $R^2=0.53$ ). The patterns of  
349 correlations between the abundance of denitrification enzymes in the two organisms and  $\text{N}_2\text{O}$   
350 (e.g., positive correlation of R12 NirK and negative correlation of 3H11 NosZ), suggested that  
351  $\text{NO}_2^-$  toxicity might be the driver of  $\text{N}_2\text{O}$  emissions in the SynCom. This hypothesis was supported  
352 by the strong correlation between both maximum concentration and cumulative (i.e., area under  
353 the curve (AUC)) levels of  $\text{NO}_2^-$  predicted by the kinetic model, and the observed final  $\text{N}_2\text{O}$  levels  
354 for each experimental context (i.e., initial  $\text{NO}_3^-$  levels and community composition). Relationships  
355 between final  $\text{N}_2\text{O}$  levels and both predicted  $\text{NO}_2^-$  metrics identified an inflection point which likely  
356 corresponds to the concentration at which  $\text{NO}_2^-$  became toxic, inhibiting process partitioning  
357 between the two organisms, which ultimately manifested in increased  $\text{N}_2\text{O}$  emissions (Fig. 5D).

358 This inflection point for N<sub>2</sub>O production occurred at ~7.5 mM NO<sub>2</sub><sup>-</sup>, which was in line with the  
359 observation that growth of 3H11 monocultures were completely inhibited at NO<sub>2</sub><sup>-</sup> concentrations  
360 >5 mM (Fig. 2A, Fig. S1).

361 Finally, we investigated factors governing NO<sub>2</sub><sup>-</sup> accumulation by using the SynCom kinetic  
362 model. We predicted the total amount of NO<sub>2</sub><sup>-</sup> produced by the SynCom across a range of initial  
363 co-culture conditions, community compositions (0.1-99.9 % 3H11) and NO<sub>3</sub><sup>-</sup> concentrations (1-30  
364 mM). This analysis revealed that NO<sub>3</sub><sup>-</sup> concentration and community composition had non-linear  
365 effects on the accumulation of NO<sub>2</sub><sup>-</sup> to a level that would result in N<sub>2</sub>O off gassing through growth  
366 inhibition of 3H11. Together our findings demonstrate that during process partitioning of  
367 denitrification, N<sub>2</sub>O production ("off gassing") is primarily driven by NO<sub>2</sub><sup>-</sup> toxicity, which manifests  
368 from a complex interplay of environmental context and community composition.

369

## 370 **DISCUSSION**

371 Our findings demonstrate how two ecologically relevant partial denitrifiers cooperate as a SynCom  
372 to perform complete denitrification (Fig.1) and how differences in mechanisms of pathway  
373 partitioning (i.e., exchange of specific metabolites) and kinetics of substrate utilization by  
374 individual members can manifest in distinct emergent phenotypes of a microbial community (Fig.  
375 2). Specifically, we discovered four key mechanisms that contributed to the overall improvement  
376 in growth dynamics of the SynCom. First, the fast rate of 3H11-mediated reduction of NO<sub>3</sub><sup>-</sup> into  
377 NO<sub>2</sub><sup>-</sup> promoted growth of R12, a finding supported by evidence that monocultures of R12 grew  
378 better on NO<sub>2</sub><sup>-</sup>, relative to NO<sub>3</sub><sup>-</sup>, with a shorter lag phase and increased growth rate (Fig. 2A and  
379 S1). Second, by consuming NO<sub>2</sub><sup>-</sup>, R12 facilitated increased biomass of 3H11, based on evidence  
380 that 3H11 was severely growth inhibited due to accumulation of NO<sub>2</sub><sup>-</sup> when grown as a  
381 monoculture (Fig. S1), but not in a SynCom context wherein NO<sub>2</sub><sup>-</sup> levels increased transiently and  
382 returned to undetectable levels (Fig. 2C and S4E). Third, the exchange of NO (instead of N<sub>2</sub>O)  
383 from R12 to 3H11 turned out to be a critical model-enabled discovery, which was supported by



384 the subsequent observation that 3H11 *norB* was significantly upregulated in the SynCom context  
385 (Fig. 3E). Finally, the finding that N<sub>2</sub>O can itself serve as a source of energy and biomass for 3H11  
386 added evidence for a fourth mechanism for synergistic improvement in specific growth rate of the  
387 SynCom (Fig. S1C). Collectively, the four mechanisms incorporated into a kinetic model explained  
388 how interplay between the two isolates improved overall growth characteristics of the SynCom  
389 (Fig. 2C).

390 While we note the natural variation in denitrification kinetics that can be achieved by  
391 environmental isolates and communities, our findings are consistent with the principle of inter-  
392 enzyme competition described by Lilja *et al.*<sup>19</sup>. The observed increase in specific growth rate as  
393 well as the increased rate of NO<sub>3</sub><sup>-</sup> reduction of the SynCom relative to R12 and 3H11 monocultures  
394 (Fig. 1E, F, S3C), suggests that inter-enzyme competition was reduced in the SynCom context.  
395 Notably, we observed that *Rhodanobacter* spp. R12 monocultures do not accumulate NO<sub>2</sub><sup>-</sup> and  
396 can grow with significantly reduced lag time and higher growth rates on NO<sub>2</sub><sup>-</sup> relative to NO<sub>3</sub><sup>-</sup> (Figs.  
397 1E and 2A). This is in contrast to the strain of *Pseudomonas stutzeri* used by Lilja *et al.*, which  
398 favors NO<sub>3</sub><sup>-</sup> reduction and thus accumulates NO<sub>2</sub><sup>-</sup><sup>19</sup>. This suggests that the R12 *Rhodanobacter*  
399 has evolved a balance between NarG and NirK, or between these enzymes and other rate limiting  
400 steps such as NO<sub>3</sub><sup>-</sup> transport, that favors reduction of NO<sub>2</sub><sup>-</sup> over NO<sub>3</sub><sup>-</sup>.

401 *Rhodanobacter* spp. dominate microbial communities within groundwater associated with  
402 low pH and high NO<sub>3</sub><sup>-</sup> in regions of the ORR, including the specific site where R12 was  
403 isolated<sup>8,23,27</sup>. Thus, the groundwater conditions at ORR might provide a selection pressure that  
404 favors the balance in NO<sub>3</sub><sup>-</sup> and NO<sub>2</sub><sup>-</sup> reduction achieved by R12 and possibly other *Rhodanobacter*  
405 spp. However, we cannot rule out the possibility that the observed improvements in specific  
406 growth rate and NO<sub>3</sub><sup>-</sup> reduction rate was due to the alleviation of NO<sub>2</sub><sup>-</sup> toxicity. Indeed, N<sub>2</sub>O off  
407 gassing by the SynCom in certain contexts (i.e., high proportion of 3H11 and high NO<sub>3</sub><sup>-</sup>) was  
408 largely attributed to the accumulation of NO<sub>2</sub><sup>-</sup> due to the increased rate of NO<sub>3</sub><sup>-</sup> reduction facilitated  
409 by 3H11, which effectively shifted the balance achieved by R12 toward NO<sub>3</sub><sup>-</sup> reduction (S3C).



410 Thus, in addition to inter-enzyme competition, natural variation in denitrification kinetics of  
411 individual organisms and toxicity of intermediates are important factors that significantly influence  
412 composition and functional interactions within a microbial community across environmental  
413 contexts.

414 Our recognition of  $\text{NO}_2^-$  accumulation as an important driver of  $\text{N}_2\text{O}$  emissions also offers  
415 mechanistic understanding to environmental studies linking  $\text{NO}_2^-$  with  $\text{N}_2\text{O}$  emissions. Maharjan  
416 and Ventura showed that  $\text{NO}_2^-$  accumulation was a major driver of  $\text{N}_2\text{O}$  emissions across different  
417 N fertilizer regimes in an agricultural setting and found that mitigation of  $\text{NO}_2^-$  accumulation  
418 reduced  $\text{N}_2\text{O}$  emissions<sup>28</sup>.  $\text{NO}_2^-$  has also been implicated in  $\text{N}_2\text{O}$  emissions in the context of  
419 wastewater treatment, where free nitrous acid ( $\text{HNO}_2$ ) has been shown to drive emissions for  
420 denitrifying-enhanced biological phosphorus removal sludge via inhibition of  $\text{N}_2\text{O}$  reduction<sup>29</sup>. In  
421 this context the authors point toward free  $\text{HNO}_2$  rather than  $\text{NO}_2^-$  itself as the source of inhibition,  
422 thus implicating pH as an important factor for the link between  $\text{NO}_2^-$  and  $\text{N}_2\text{O}$  emissions in  
423 denitrifying communities. Therefore, while the inhibitory effects of  $\text{NO}_2^-$  are well characterized and  
424 have been associated with  $\text{N}_2\text{O}$  emission in other contexts, here using the 3H11-R12 SynCom  
425 we have uncovered how factors, such as  $\text{NO}_3^-$  concentration, community composition, and  
426 substrate utilization kinetics, modulate partitioning of denitrification across a microbial community  
427 to drive  $\text{NO}_2^-$  accumulation and  $\text{N}_2\text{O}$  emissions.

428 While the kinetic model provided a mechanistic explanation for improved growth  
429 characteristics of the SynCom, it also suggested that functional interactions between the two  
430 organisms extended beyond the core process of denitrification. Indeed, gene expression and  
431 proteomics analyses uncovered widespread complementary physiological changes across both  
432 organisms when they were grown as a SynCom (Figs. 3 and 4). Correlated protein level changes  
433 of enzymes of BCAA and histidine biosynthesis in 3H11 with corresponding changes in  $\text{NO}_3^-$  and  
434  $\text{N}_2\text{O}$  levels suggested that upregulation of these pathways was likely governed by the availability  
435 of NO produced by R12. Upregulation of BCAA biosynthesis may contribute towards maintaining

436 intracellular redox balance for 3H11 and R12, as reported previously in the purple non sulfur  
437 bacterium *Rhodospirillum rubrum* growing under photoheterotrophic conditions<sup>30,31</sup>. Additionally,  
438 upregulation of leucine export by 3H11 and BCAA degradation genes by R12 (e.g., leucine  
439 dehydrogenase, and propionyl-CoA carboxylase) suggested exchange of leucine from 3H11 to  
440 R12. Similar trends in protein level changes were observed which suggested that methionine  
441 biosynthesis by 3H11 was likely leveraged by R12 through upregulation of uptake and  
442 degradation pathways. These observations collectively suggest that when growing in a SynCom  
443 context, R12 derives some of its nutritional needs from 3H11, which may be facilitated by  
444 coordinated upregulation of flagellar genes<sup>32</sup>. Consistent with these findings, simulations based  
445 on the stoichiometric constraints-based metabolic model of the SynCom predicted that exchange  
446 of intermediates of central C metabolism and amino acids could contribute up to 10% increase in  
447 overall biomass of the SynCom (Fig. S3C). In sum, multiomics profiling and model simulations  
448 suggested that synergistic improvement in growth characteristics of the SynCom was likely an  
449 outcome of process partitioning across both denitrification as well as other metabolic pathways,  
450 explaining why metabolite exchange is ubiquitous across natural microbial communities<sup>33</sup>. Future  
451 work should integrate kinetics of denitrification into metabolic models of monocultures and  
452 SynComs to better understand the dynamics of community metabolism and how it changes with  
453 growth context<sup>34,35</sup>. This may allow us to better understand how changes in pH, temperature,  
454 dissolved oxygen, and other environmentally relevant factors influence cellular physiology,  
455 community cooperativity, and the production of N<sub>2</sub>O in natural ecosystems.

456

## 457 **METHODS**

### 458 **ORR and IMG isolate denitrification pathway composition analysis**

459 Annotated ORR isolate genomes were obtained from KBase  
460 (<https://narrative.kbase.us/narrative/55476>). Additional soil isolate genomes were obtained from  
461 the Joint Genome Institute's (JGI's) Integrated Microbial Genomes & Microbiomes (IMG/M)

462 database were. The presence of the denitrification genes Nar/Nap, Nir, Nor, and Nos were  
463 characterized using KEGG (Kyoto Encyclopedia of Genes and Genomes) Orthology (KO) terms.  
464 Isolates with one or more denitrification genes were considered across both datasets.  
465 Denitrification pathway composition of the isolates was tabulated using gene presence/absence.

466

#### 467 **ORR groundwater genus abundance and chemistry analysis**

468 Groundwater OTU abundance, taxonomy, and chemical data previously described by Smith et al.  
469 2015 were obtained and re-analyzed<sup>20</sup>. OTU relative abundances were summarized at the genus  
470 level and log<sub>10</sub> transformed. Chemical concentrations were converted to mM and log<sub>10</sub>  
471 transformed where appropriate. Genus presence/absence was determined by examining the  
472 overall distribution of log<sub>10</sub> transformed genus relative abundances and setting a threshold.  
473 Genera with a log<sub>10</sub> abundance > -6 (e.g., relative abundance > 10<sup>-6</sup>) were considered present.  
474 This threshold separated the peaks of the bi-model log<sub>10</sub> genus relative abundance distribution.  
475 Statistical analyses, including linear regression, and significance testing, were performed using  
476 tools from the python Scikit-Learn, SciPy, and NumPy packages<sup>36-38</sup>.

477

#### 478 **Isolate and SynCom and metabolic models**

479 The genome-scale metabolic models (GEMs) used in this study were reconstructed using the  
480 ModelSEED v2<sup>39</sup> app in The Department of Energy (DOE)'s Systems Biology Knowledgebase  
481 (KBase)<sup>40</sup>. Flux balance analysis (FBA)<sup>41</sup> was used to perform growth simulations for individual  
482 models for R12 and 3H11 and species interactions in the SynCom. The biomass objective  
483 function of the individual models was adjusted as needed to represent the abundance of each  
484 organism in the SynCom. When simulating community interactions, we limited the exchange of  
485 compounds between organisms to the compounds present in the *in silico* growth media.

486

#### 487 **Strains and medium preparation**

488 *Rhodanobacter* sp. *FW510-R12* (R12) and *Acidovorax* sp. *GW101-3H11* (3H11) were both  
489 isolated from contaminated wells at the Oak Ridge Integrated Field Research Center (ORR) as  
490 reported in Hemme et al. 2016 and Price et al. 2018, respectively <sup>42,43</sup>. NO<sub>3</sub> reduction growth  
491 studies were performed in Balch tubes (10-mL culture volume) at 30°C in a defined minimal  
492 medium as described previously containing 20 mM sodium acetate and 10 mM sodium NO<sub>3</sub> at pH  
493 7.2 with a 80:20 N<sub>2</sub>-CO<sub>2</sub> headspace <sup>9</sup>. A small amount of yeast extract (0.1 g/L) was added to the  
494 medium before autoclaving to support growth of 3H11 and R12. For experiments that  
495 characterized growth on variable amounts of NO<sub>3</sub>, NO<sub>2</sub><sup>-</sup>, and N<sub>2</sub>O media was first prepared in  
496 serum bottles. Vitamin and phosphate solutions were added to serum bottle media and then  
497 aliquoted into N<sub>2</sub>-CO<sub>2</sub> flushed and pressurized Balch tubes (10 mL each, ~1 psi overpressure).  
498 For experiments characterizing growth of 3H11 on N<sub>2</sub>O, injections of pure N<sub>2</sub>O (0-2.59 mL at 14.5  
499 psi overpressure) into Balch tubes were performed to achieve desired concentrations (0-20 mM).

500

#### 501 **Preparation of R12 and 3H11 co-cultures and pure cultures**

502 From freezer stocks, cells were plated on R2A agar and incubated at 30 °C. Following growth of  
503 isolated colonies, colonies were picked and used for inoculation of R2A liquid medium cultures.  
504 Liquid R2A cultures were incubated at 30°C with shaking at 100 rpm. Following overnight growth,  
505 optical density at 600 nm (OD<sub>600</sub>) was measured and cell density was normalized to using R2A  
506 liquid medium if necessary. With starting cultures at ~0.5 OD<sub>600</sub> units, 100 uL of *Rhodanobacter*  
507 sp. *FW510-R12* (R12) and 100 uL of *Acidovorax* *GW101-3H11* (3H11) were inoculated into a NO<sub>3</sub>  
508 reduction medium to establish co-cultures. For pure cultures 200 uL of either strain was utilized  
509 to normalize starting cell densities. Initial cell densities were normalized to ~0.01 OD<sub>600</sub> units  
510 across all experiments and conditions. Biological replicates varied from 3-8 across experiments.  
511 For all experiments cell concentration was monitored in Balch tubes with periodic measurements

512 of the optical density. Initial experiments (Figs. 1E, 2C, and 3A) measured optical density 1-3  
513 times daily using a Spectronic 200 spectrophotometer (Thermo Fisher). Later experiments (Figs.  
514 2A, 4, and 5) leverage an automated OD600 measurement with shaking at 100 rpm previously  
515 described <sup>44</sup>. In this context, measurements were taken every 5-30 minutes and instrument  
516 voltage was converted to OD600 by calibration. Sample OD600 was measured before and after  
517 each experiment to calibrate individual sample voltage ranges.

518

### 519 **Measurement of N species**

520 NO<sub>3</sub>, NO<sub>2</sub><sup>-</sup>, and acetate concentrations were analyzed using the ion-chromatography Dionex ICS-  
521 5000 system with the IonPac® ICE-AS6 column (Thermo Fisher). Medium N<sub>2</sub>O concentration was  
522 quantified by sampling 200 uL medium into GC vials containing 200 uL 8% PFA in 1x PBS (to  
523 achieve a final concentration of 4% PFA). Vials were allowed to equilibrate and headspace N<sub>2</sub>O  
524 was measured using a Shimadzu GC-2014 with an AOC-6000 autosampler. Vials were  
525 equilibrated for 60 seconds in a 35 °C oven with shaking, and then a 1000 µL sample was  
526 aspirated from the headspace of the vial with a syringe preheated to 100 °C. Run conditions were  
527 as follows: inlet temperature 325 °C, oven temperature 80 °C, column flow rate 20 mL/min, TCD  
528 temperature: 150 °C, current: 80 mA, ECD temperature: 325 °C, current: 2 nA, run time: 20  
529 minutes, carrier gas: argon, makeup gas: N. GC vial headspace N<sub>2</sub>O ppm values were converted  
530 to culture vessel liquid N<sub>2</sub>O concentration (mM) using Henry's law. Total culture vessel N<sub>2</sub>O  
531 concentration reported was obtained using mass balance and assuming equilibrium.

532

### 533 **Monoculture and SynCom kinetic models**

534 Kinetic models were developed using a modified Monod framework by integrating Logistic  
535 representation of carrying capacity into equations describing growth kinetics as a function of  
536 metabolite concentrations <sup>24</sup>. Equations in the model were of the form  $\frac{1}{X} \frac{dX}{dt} = \mu =$   
537  $\mu_m \left( \frac{S}{S+K_m} \right) \left( 1 - \frac{X}{X_m} \right)$ , where  $X$  is cell population size,  $S$  is substrate concentration,  $K_m$  and  $\mu_m$  are  
538 substrate specific Monod kinetic parameters, and  $X_m$  is cell specific capacity. Substrate specific  
539 rates were assumed to be additive (e.g.,  $\mu_{R12} = \mu_{NO_3} + \mu_{NO_2} + \mu_{NO}$ ). Equations were  
540 parameterized using maximum growth rates as a function of substrate, half velocity constants,  
541 and carrying capacities extracted from species specific growth data using Logistic and Monod fits.  
542 To represent carrying capacity as a function of substrate concentration, linear models were  
543 developed to capture relationships between  $NO_3^-$ ,  $NO_2^-$ ,  $N_2O$  concentration and maximum OD600.  
544 Carrying capacity models were of the form  $OD_{R12} = \beta_{NO_3}X_{NO_3} + \beta_{NO_2}X_{NO_2}$  and  $OD_{3H11} =$   
545  $\beta_{NO_3}X_{NO_3} + \beta_{NO_2}X_{NO_2} + \beta_{N_2O}X_{N_2O}$  for R12 and 3H11 respectively. Linear model parameters were  
546 estimated by OLS fit of monoculture maximum OD600 and initial substrate concentration data.  
547 Kinetic model simulations were performed using the python package Tellurium <sup>45</sup>. Statistical  
548 analyses, including linear regression, and significance testing, were performed using tools from  
549 the python Scikit-Learn, SciPy, and NumPy packages <sup>36-38</sup>.

550

## 551 **Transcriptomics profiling and analysis**

552 **(i) Sample collection and sequencing.** R12-3H11 co-cultures and pure cultures were harvested  
553 for transcriptome profiling in biological triplicates across time points (Fig. 3). Cell pellets were  
554 harvested by centrifugation at 4,000 × g and flash frozen in liquid N<sub>2</sub>. Total RNA was extracted  
555 using the MasterPure™ Complete DNA and RNA Purification Kit from Epicentre. All samples were  
556 treated with Invitrogen TURBO DNA-Free kit to remove DNA contamination. The Illumina  
557 Stranded Total RNA Prep Ligation with Ribo-Zero Plus (Illumina) was used for rRNA depletion

558 and library preparation. Sequencing was performed using the NextSeq 500 platform (2 by 75 bp,  
559 Illumina) with 10 to 15 million reads per sample.

560 **(ii) Read processing.** RNA sequencing reads were analyzed with FastQC according to Illumina's  
561 default quality filtering process and then trimmed using base quality scores by Trimmomatic<sup>46,47</sup>.  
562 A quality score of 20 was used for read trimming and quality filtering. The abundance of 3H11  
563 and R12 specific transcripts were quantified using the concatenated genomes of *Rhodanobacter*  
564 *sp. FW510-R12* (NCBI BioProject PRJNA255897) and *Acidovorax sp. GW101-3H11* (NCBI  
565 BioProject PRJNA314893) and Kallisto<sup>48</sup>.

566 **(iii) Differential expression analysis, clustering, and functional enrichment.** Normalized  
567 expression data and differentially expressed genes (DEGs) were obtained from DESeq2<sup>25</sup> using  
568 pairwise comparisons of growth phase matched monoculture and SynCom transcriptomes.  
569 Genes were reported as dysregulated if the log<sub>2</sub> fold change magnitude  $\geq 2$  and significant  
570 ( $p < 0.05$  and Benjamini-Hochberg false discovery rate  $q < 0.01$ ). DEGs were clustered using the  
571 scikit-learn implementation of the k-means algorithm<sup>36</sup>. Statistical analyses, including principal-  
572 component analysis (PCA), significance testing, and functional enrichment were performed using  
573 tools from the Python Scikit-learn, SciPy, and NumPy packages<sup>36-38</sup>. The significance of KEGG  
574 subpathway term enrichment among k-means clusters was assessed by comparing term  
575 frequencies within each cluster to their frequencies in the genome using the hypergeometric test  
576<sup>49</sup>. Significantly enriched terms ( $p < 0.05$  and Benjamini-Hochberg false-discovery rate  $q < 0.01$ )  
577 were reported<sup>50</sup>. KEGG pathway level expression comparisons were assessed using DESeq2  
578 normalized gene expression Z-scores and Welch's t-test.

579

580 **Proteomics profiling and analysis**



581 **(i) Sample collection and peptide quantification.** R12-3H11 co-cultures and pure cultures were  
582 harvested prior to inoculation and at the end of growth (stationary phase) for proteomics profiling  
583 using 8 biological replicates (Fig. 4, S4). Cell pellets were harvested by centrifugation at 4,000 ×  
584 g and flash frozen in liquid N<sub>2</sub> and stored at -80 °C until further processing. Protein was extracted  
585 from cell pellets and tryptic peptides were prepared by following an established proteomic sample  
586 preparation protocol<sup>51</sup>. Briefly, cell pellets were resuspended in Qiagen P2 Lysis Buffer (Qiagen,  
587 Germany) to promote cell lysis. Proteins were precipitated with addition of 1 mM NaCl and 4 x vol  
588 acetone, followed by two additional washes with 80% acetone in water. The recovered protein  
589 pellet was homogenized by mixing with 100 mM ammonium bicarbonate in 20% methanol. Protein  
590 concentration was determined by the DC protein assay (BioRad, USA). Protein reduction was  
591 accomplished using 5 mM tris 2-(carboxyethyl)phosphine (TCEP) for 30 min at room temperature,  
592 and alkylation was performed with 10 mM iodoacetamide (IAM; final concentration) for 30 min at  
593 room temperature in the dark. Overnight digestion with trypsin was accomplished with a 1:50  
594 trypsin:total protein ratio. The resulting peptide samples were analyzed on an Agilent 1290  
595 UHPLC system coupled to a Thermo Scientific Orbitrap Exploris 480 mass spectrometer for  
596 discovery proteomics<sup>52</sup>. Briefly, peptide samples were loaded onto an Ascentis® ES-C18 Column  
597 (Sigma–Aldrich, USA) and were eluted from the column by using a 10 minute gradient from 98%  
598 solvent A (0.1 % FA in water) and 2% solvent B (0.1% FA in ACN) to 65% solvent A and 35%  
599 solvent B. Eluting peptides were introduced to the mass spectrometer operating in positive-ion  
600 mode and were measured in data-independent acquisition (DIA) mode with a duty cycle of 3  
601 survey scans from m/z 380 to m/z 985 and 45 MS2 scans with precursor isolation width of 13.5  
602 m/z to cover the mass range. DIA raw data files were analyzed by an integrated software suite  
603 DIA-NN<sup>53</sup>. The database used in the DIA-NN search (library-free mode) leveraged the combined  
604 3H11 and R12 annotated proteome FASTA sequences plus the protein sequences of common  
605 proteomic contaminants. DIA-NN determines mass tolerances automatically based on first pass  
606 analysis of the samples with automated determination of optimal mass accuracies. The retention



607 time extraction window was determined individually for all MS runs analyzed via the automated  
608 optimization procedure implemented in DIA-NN. Protein inference was enabled, and the  
609 quantification strategy was set to Robust LC=High Accuracy. Output main DIA-NN reports were  
610 filtered with a global FDR=0.01 on both the precursor level and protein group level.

611 **(ii) Species relative abundance estimates.** The quantification of species biomass contributions  
612 in the community was based on the input amounts of the sum of peptide ion intensities using only  
613 unique peptides from each species. This method has been benchmarked previously as one of the  
614 three proteomic quantification methods for accurate estimation of species-level biomass  
615 contribution in microbial communities <sup>26</sup>.

616 **(iii) Statistical analyses.** Statistical analyses, including PCA, linear regression, and significance  
617 testing, were performed using tools from the Python Scikit-learn, SciPy, and NumPy packages <sup>36–</sup>  
618 <sup>38</sup>. The significance of individual protein abundance changes was assessed by correlating protein  
619 abundance with initial NO<sub>3</sub><sup>-</sup> levels, inoculum 3H11 percentage, and steady state N<sub>2</sub>O levels.  
620 Pearson *r* values were used to identify significant hits, proteins with  $r \leq 0.5$  or  $r \geq 0.5$  and  $q < 0.05$   
621 (Benjamini-Hochberg false-discovery rate correction) were considered <sup>50</sup>. KEGG subpathway  
622 level protein abundance correlations with steady state N<sub>2</sub>O levels were performed using data  
623 randomly split into training (70% of samples) and testing (30% samples) sets. KEGG subpathway  
624 terms were used to identify groups of proteins and multi-linear regression was performed with the  
625 training set. Model accuracy was then assessed using the test set and out of sample R<sup>2</sup> was  
626 reported.

## DATA AVAILABILITY

Data summary tables (e.g., growth and metabolite concentration data, processed transcriptomics and proteomics data, and metabolic model outputs) and code used for data analysis and figure generation can be found at Github at [https://github.com/baliqa-lab/3H11\\_R12\\_SynCom](https://github.com/baliqa-lab/3H11_R12_SynCom). Raw sequencing data used for the transcriptomics data was uploaded to NCBI GEO repository with the dataset identifier GSE272493. The generated mass spectrometry proteomics data have been deposited to the ProteomeXchange Consortium via the PRIDE partner repository with the dataset identifier PXD051979<sup>54</sup>.

## ACKNOWLEDGMENTS

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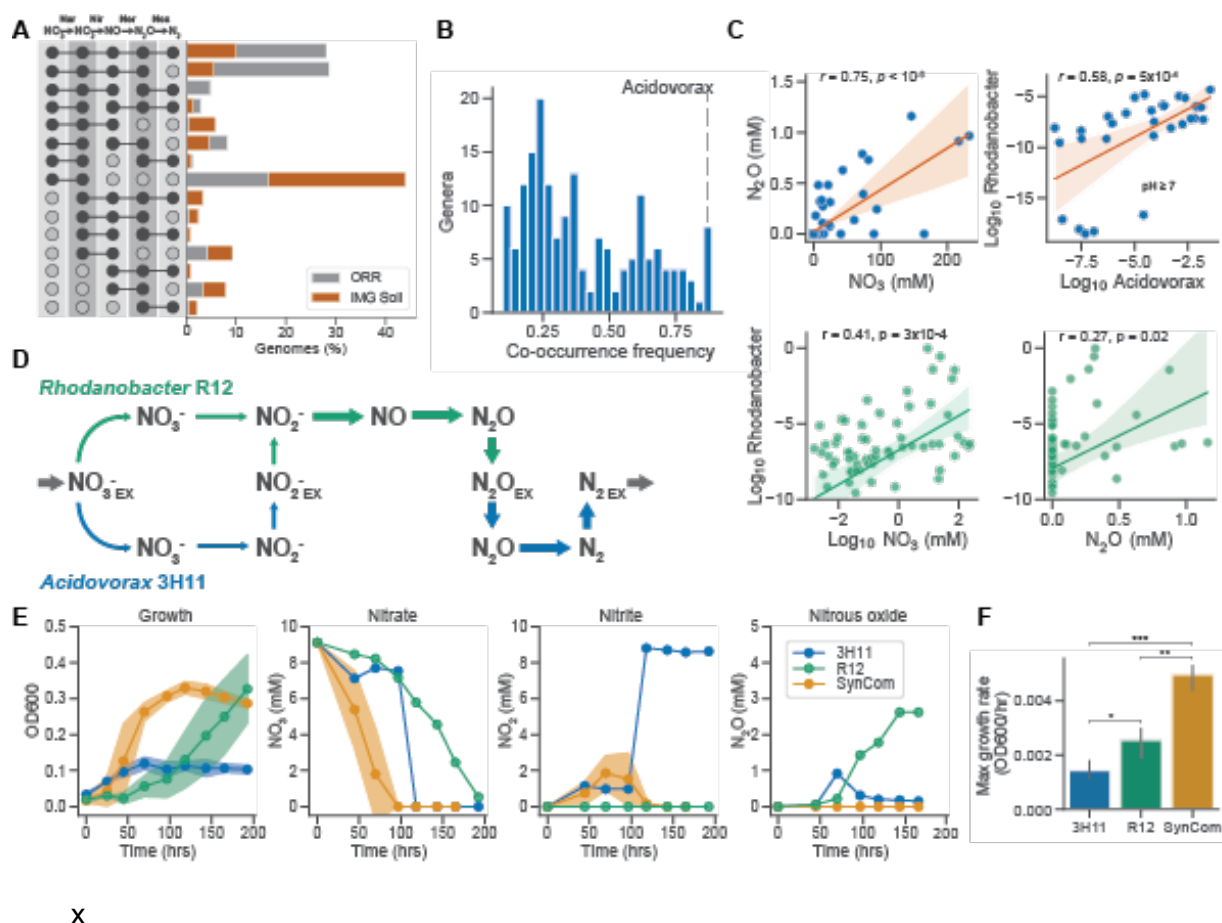
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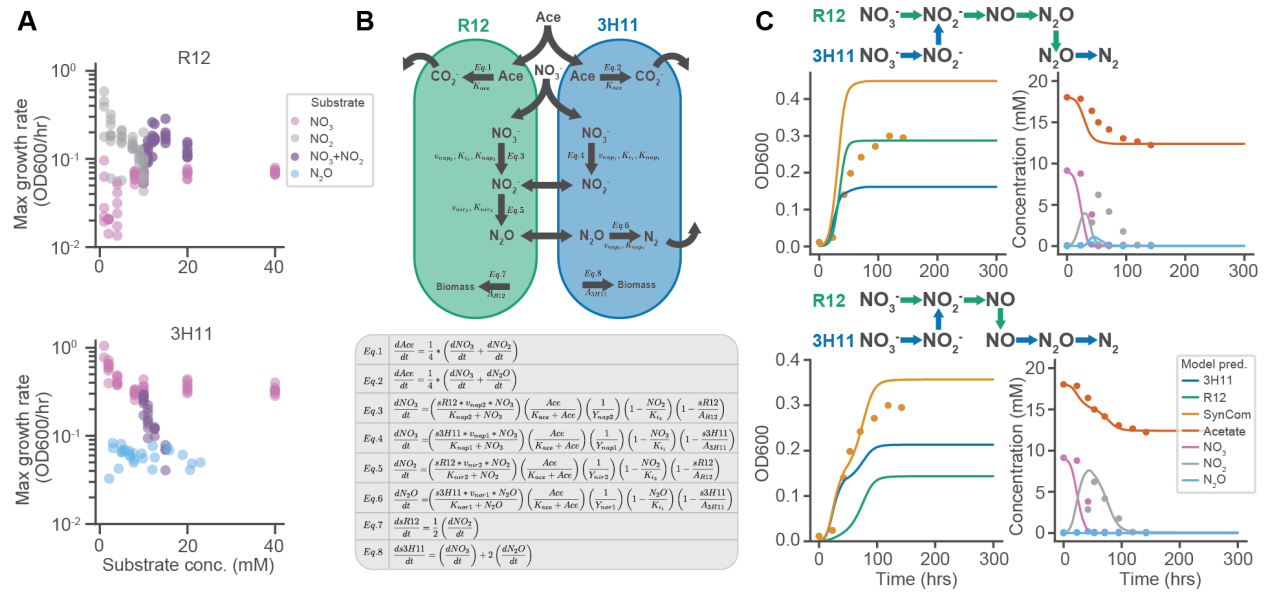


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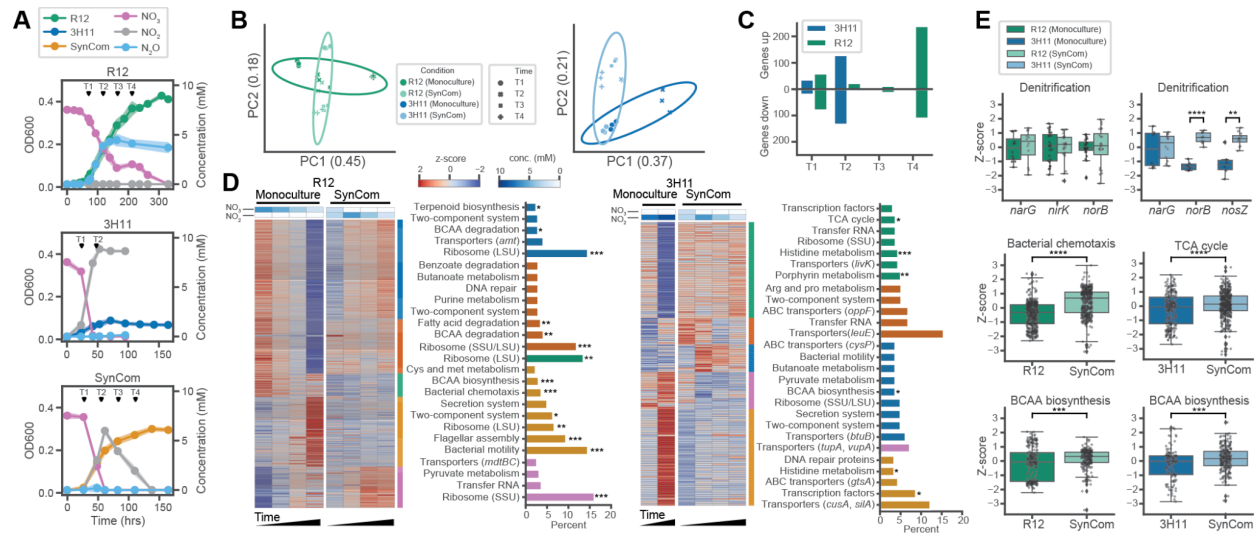
**Figure 1. Field observations and metabolic modeling predict exchange of intermediates in a denitrifying synthetic community.** (A) Distribution of denitrification pathway compositions among ORR field and IMG soil isolates. (B) Genus level co-occurrence frequency of *Rhodanobacter* and other genera across ORR groundwater samples. (C) Association between groundwater chemistries and genus abundance. Pearson correlation  $r$  and  $p$  values are listed in each subplot. Linear fits to data using OLS are displayed as well as 95% confidence intervals (D) Denitrification flux patterns predicted by genome scale metabolic model of R12 and 3H11 SynCom assuming equal abundances. (E) Growth characteristics of R12, 3H11 and R12-3H11 SynCom. Shading around trend lines represents standard deviation of samples and points represent averages. (F) Comparison of monoculture and SynCom maximum growth rates. Bars



indicate comparisons for which differences were significant using Welch's t-test. \*,  $p < 0.05$ ; \*\*,  $p < 0.01$ ; \*\*\*,  $p < 0.001$ .

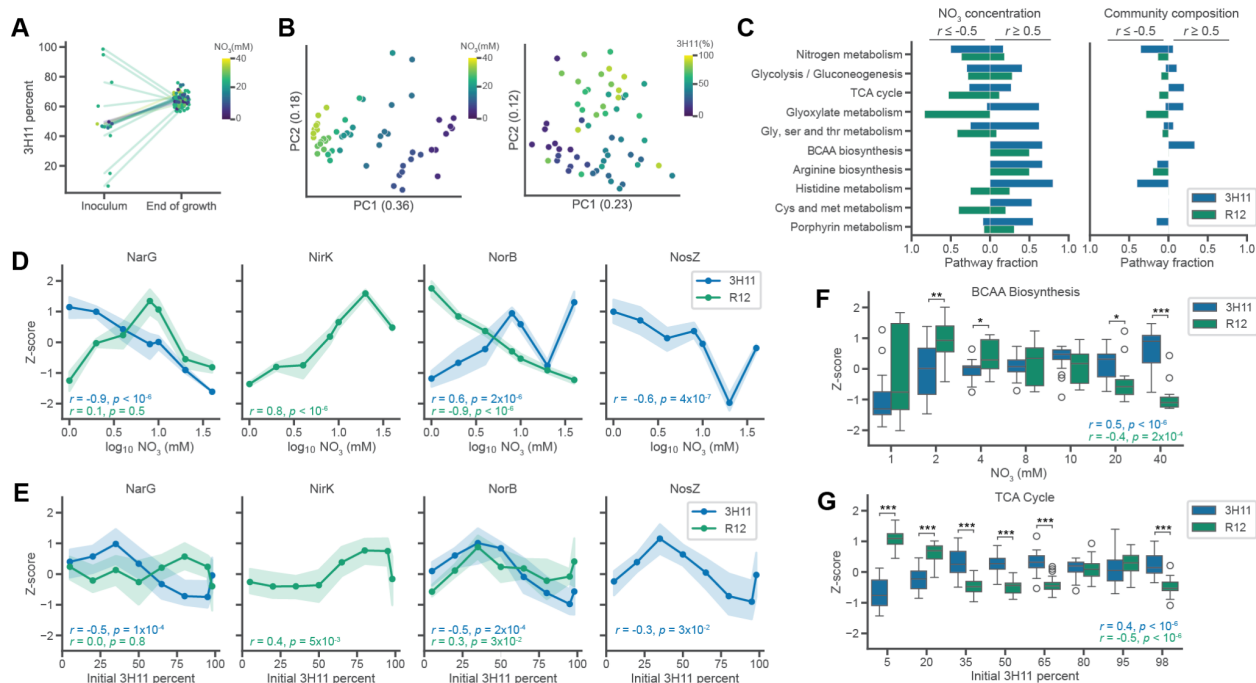


**Figure 2. Exchange of pathway intermediates drives growth enhancement in R12-3H11 SynCom.** (A) Maximum growth rates of R12 and 3H11 across variation in substrate concentrations. For conditions where NO<sub>3</sub><sup>-</sup> and NO<sub>2</sub><sup>-</sup> were present, 10 mM NO<sub>3</sub><sup>-</sup> was used and NO<sub>2</sub><sup>-</sup> was varied. (B) Schematic of R12-3H11 kinetic model and associated differential equations. Model represents growth resulting from oxidation of acetate and reduction of NO<sub>3</sub><sup>-</sup>, NO<sub>2</sub><sup>-</sup>, and N<sub>2</sub>O via denitrification. Given challenges of measuring NO kinetics, NO reduction is not explicitly represented in the model, instead NO<sub>2</sub><sup>-</sup> is reduced to N<sub>2</sub>O (C) Two SynCom intermediate exchange scenarios and associated dynamics predicted by kinetic models displayed with empirical data. Model predictions are represented using solid lines, data are averages across samples and are represented using circular points.



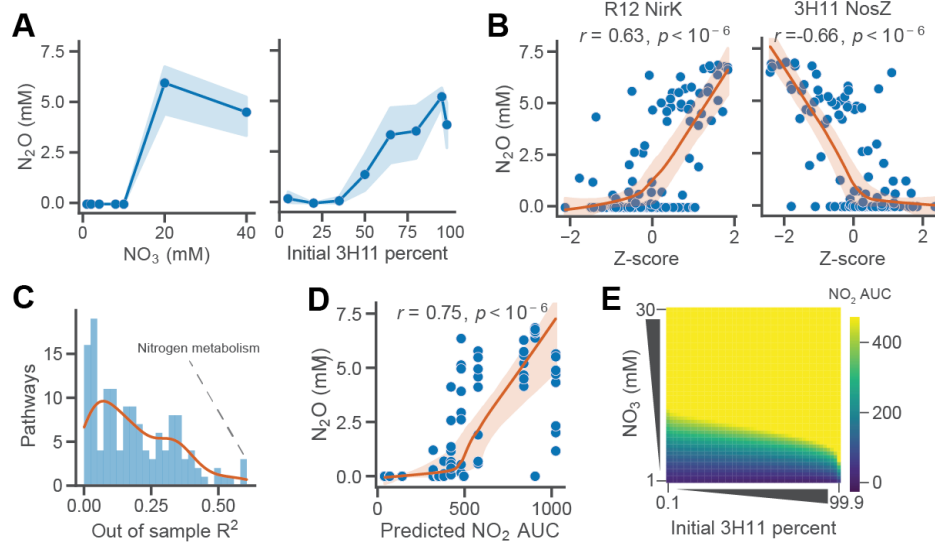
**Figure 3. Community context drives changes in the global transcriptional states of 3H11 and R12.** (A) Growth characteristics of R12, 3H11, the SynCom, and associated transcriptomics sampling points. Black triangles mark condition specific time points (T1-T4). Shading around trend lines represents standard deviations and points represent averages. (B) PCA of R12 and 3H11 gene expression across monoculture and co-culture growth contexts. DESeq2 log normalized gene expression Z-scores used. Z-scores computed across growth contexts for R12 and 3H11. (C) R12 and 3H11 specific DEG counts for comparisons across time and growth contexts. (D) K-means clustered R12 and 3H11 specific DEGs using k=5. Expression Z-scores are displayed across time and growth contexts. Associated  $\text{NO}_3^-$  and  $\text{NO}_2^-$  concentrations are displayed above each heatmap. To the right of each heat map, the most abundant KEGG subpathway terms are displayed for each cluster. Bars are colored by the cluster they represent. For ribosome terms, the most abundant subunit is noted. Large subunit (LSU) or small subunit (SSU). For ABC transporter and transporter terms, a representative gene in each cluster is noted (See Supplementary Table 1 for full list of DEGs and associated clusters). Subpathway terms are noted which were significantly enriched in each cluster using the hypergeometric test. \*,  $p < 0.05$ ; \*\*,  $p < 0.01$ ; \*\*\*,  $p < 0.001$  (E). Boxplots of gene and pathway expression Z-scores. Bars indicate

comparisons for which differences were significant using Welch's t-test. \*,  $p < 0.05$ ; \*\*,  $p < 0.01$ ; \*\*\*,  $p < 0.001$

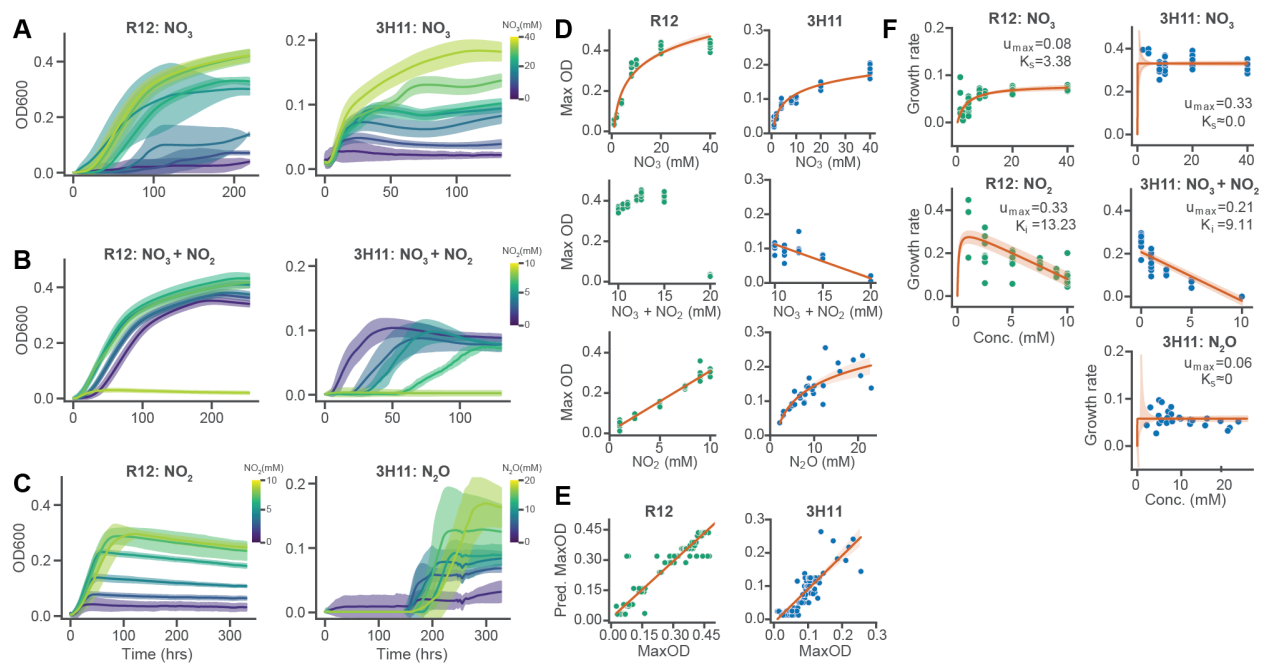


**Figure 4. Variation in  $\text{NO}_3^-$  and community composition support the dominant role of 3H11 in the SynCom.** (A) 3H11 SynCom percentage across variation in initial community composition and  $\text{NO}_3^-$  concentration. Community composition measured before inoculation and at the end of growth. Color indicates  $\text{NO}_3^-$  concentration used and lines connect initial and end points of individual cultures. (B) PCA of SynCom end-point proteome Z-scores across variation in initial community composition and  $\text{NO}_3^-$  concentration. Color used to represent growth conditions. (C) Fraction of detected 3H11 and R12 specific pathway proteins significantly correlated with  $\text{NO}_3^-$  and community composition. (D) End-point 3H11 and R12 denitrification protein Z-scores as a function of initial  $\text{NO}_3^-$  concentration. (E) End-point 3H11 and R12 denitrification protein Z-scores as a function of initial community composition. Points represent an average of eight replicates, error bands display standard deviation. 3H11 and R12 specific Pearson  $r$  and  $p$  values are displayed on each subplot. (F) Boxplots of BCAA biosynthesis protein Z-scores across  $\text{NO}_3^-$  concentration. (G) Boxplots of TCA cycle protein Z-scores across community composition. Bars

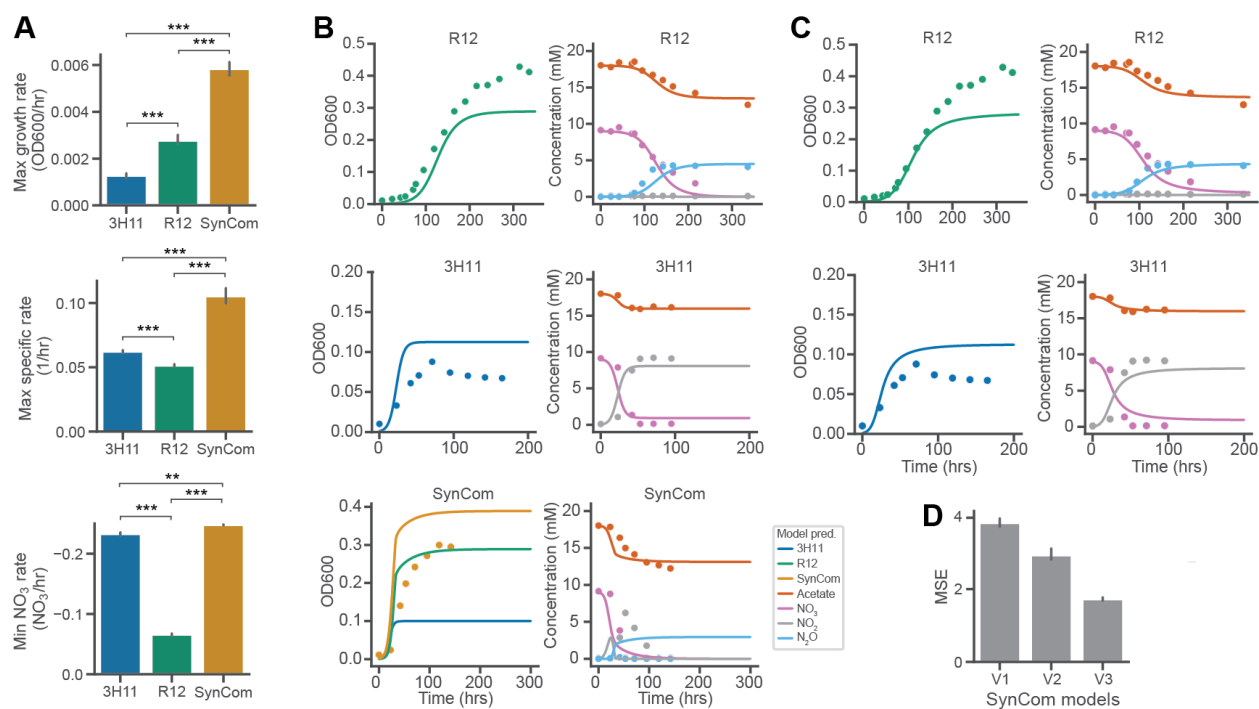
indicate comparisons for which differences were significant using Welch's t-test. \*,  $p < 0.05$ ; \*\*,  $p < 0.01$ ; \*\*\*,  $p < 0.001$ .



**Figure 5.  $\text{NO}_2^-$  inhibition drives  $\text{N}_2\text{O}$  accumulation in the SynCom.** (A) End-point SynCom  $\text{N}_2\text{O}$  concentration as a function of initial  $\text{NO}_3^-$  concentration and community composition. Points represent an average of eight replicates, error bands display standard deviation. (B) End-point SynCom  $\text{N}_2\text{O}$  concentration as a function of denitrification protein Z-scores across conditions. Lowest trendline and 95% confidence interval displayed. Pearson  $r$  and  $p$  values displayed. (C) Distribution of predicted  $\text{N}_2\text{O}$  out of sample  $R^2$  using pathway level multi-linear regression.  $R^2$  of N metabolism proteins noted. (D) End-point SynCom  $\text{N}_2\text{O}$  concentration as a function of predicted SynCom  $\text{NO}_2^-$  AUC concentration across conditions. Lowest trendline and OLS linear fit displayed with respective 95% confidence intervals. Pearson  $r$  and  $p$  values displayed. (E) Heatmap of predicted SynCom  $\text{NO}_2^-$  AUC across variation in initial  $\text{NO}_3^-$  concentration and community composition.

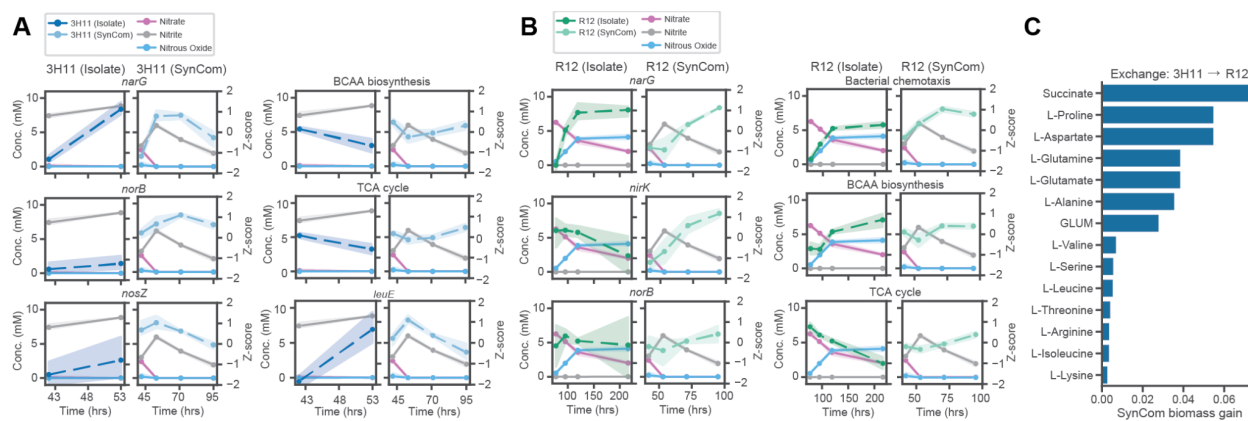


**Supplemental figure 1. Growth characteristics of 3H11 and R12 across variation in substrate concentration.** (A) Growth of R12 and 3H11 across variation in initial NO<sub>3</sub> concentration. Trendlines display average trajectories across eight replicates and error bands display standard deviations. (B) Growth of R12 and 3H11 across variation in initial NO<sub>2</sub> concentration supplemented with 10 mM NO<sub>3</sub>. (C) Growth of R12 and 3H11 across variation in initial NO<sub>2</sub>- and N<sub>2</sub>O concentration respectively. Trendlines display average trajectories across eight replicates and error bands display standard deviations. Trendlines are colored by respective substrate concentrations. (D) Maximum OD600 achieved by 3H11 and R12 across growth conditions as a function of substrate concentration. OLS linear or logarithmic fits to data are displayed with 95% confidence intervals where appropriate. (E) Predicted maximum OD600 vs measured maximum OD600 using multi-linear fit to linear portions of substrate vs maximum OD600 data. (F) Maximum growth rate as a function of substrate concentration across conditions for 3H11 and R12. Fits of Monod substrate kinetics or Monod product inhibition kinetics displayed with 95% confidence intervals and associated parameters.



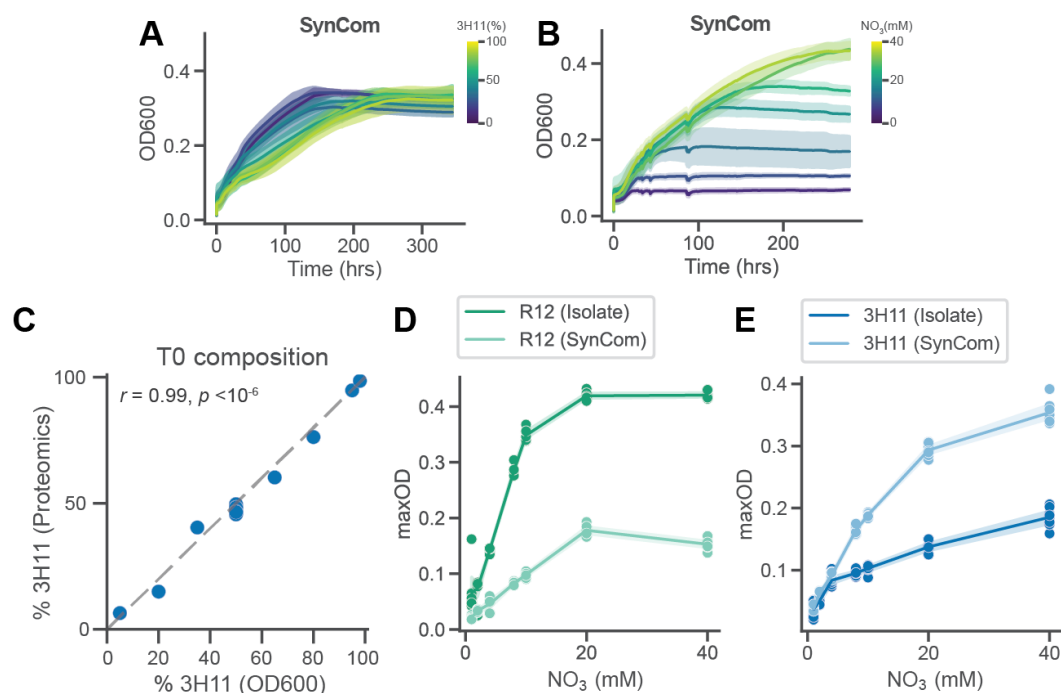
### Supplemental figure 2. Kinetic modeling of individual monocultures and SynCom growth

**characteristics.** (A) Comparison of monoculture and SynCom maximum growth rates, maximum specific rates and minimum NO<sub>3</sub><sup>-</sup> rate from the transcriptomics growth data. (B) Predicted growth dynamics and associated data for initial kinetic model formulations for R12, 3H11 and the SynCom. (C) Updated growth dynamics predictions with models which included NO<sub>3</sub><sup>-</sup> inhibition. Model predictions are represented using solid lines, data are averages across triplicate samples and are represented using circular points. SynCom models assume NO<sub>2</sub><sup>-</sup> and N<sub>2</sub>O are exchanged. (D) Comparison of mean standard error (MSE) for three iterations of SynCom kinetic models. MSE computed across both growth and metabolite data. Bars indicate comparisons for which differences were significant using Welch's t-test. \*,  $p < 0.05$ ; \*\*,  $p < 0.01$ ; \*\*\*,  $p < 0.001$ .



**Supplemental figure 3. Time resolved expression dynamics of denitrification genes and select pathways. (A)** 3H11 Expression dynamics for specific genes and pathways. **(B)** R12 Expression dynamics for specific genes and pathways. Points and trend lines represent averages of triplicate samples in the case of genes and averages across genes and samples in the case of pathways. Error bands display standard deviations. **(C)** Simulations with the constrains based metabolic model predict modest increase in biomass of the SyCom through exchange of amino acids and intermediates of central carbon metabolism.





**Supplemental figure 4. SynCom growth characteristics across variation in initial community composition and NO<sub>3</sub> concentration.** (A) Growth of the SynCpm across variation in initial community composition (3H11 %). Trendlines display average trajectories across eight replicates and error bands display standard deviations. Trendlines are colored by 3H11 %. (B) Growth of the SynCpm across variation in initial NO<sub>3</sub> concentrations. Trendlines display average trajectories across eight replicates and error bands display standard deviations. Trendlines are colored by NO<sub>3</sub> concentration. (C) Correlation between inoculum 3H11 % estimated by OD600 and proteomics. Dashed line indicates a 1:1 relationship. (D) Comparison of maximum OD600 achieved by R12 in monoculture vs R12 in the SynCom as a function of NO<sub>3</sub> concentration. (E) Comparison of maximum OD600 achieved by 3H11 in monoculture vs 3H11 in the SynCom as a function of NO<sub>3</sub> concentration.